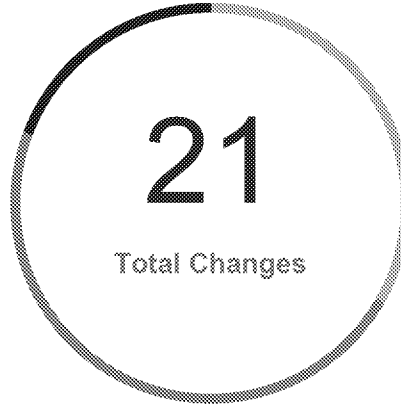





Summary



-  Shows Replacements
-  Shows Insertions
-  Shows Deletions

Appendix B

Development and Use of Risk-Based Concentrations to Select an Area for Remedial Action

This appendix describes the development and use of site-specific Risk-Based Concentrations (RBCs) to identify the area where remedial action is appropriate for soil at the Rolling Knolls Landfill Superfund Site located in the Township of Chatham, Morris County, New Jersey (the “Site”). To support this evaluation, an iterative approach was used to compare exposure point concentrations (EPCs) of polychlorinated biphenyls (PCBs), the primary risk driver and only constituent with an individual constituent hazard index above one (1), outside the proposed area to be remediated to the Site-specific RBCs.

BACKGROUND INFORMATION

Site Features

The Site is a former municipal landfill in use from the 1930s to 1968. It consists of approximately 140 acres of landfill, with an adjacent 30-acre area west of the landfill that has debris scattered on the surface, but no buried waste (known as the Surface Debris Area). The Site features are shown on Figure B-1. Most of the landfill and the Surface Debris Area are privately owned. Approximately 35 acres of the landfill are on the Great Swamp National Wildlife Refuge (GSNWR).

The Site also includes a shooting range and ball field located north of the landfill.

Current Uses

Currently, a Baseball Field and a Shooting Range are located north of the landfill and are used occasionally for recreation. A small building known as the Hunt Club is located in the Surface Debris Area near the western boundary of the landfill; it is generally unoccupied but is used occasionally for social gatherings. Two areas of the Site (Landscape Areas 1 and 2) are leased to landscaping firms for the storage of trucks and equipment. An area of the Site is also used by South Orange Disposal for the storage of roll offs. ♻️

Previous Risk Assessments

The Baseline Human Health Risk Assessment (BHHRA) prepared by CDM Federal Programs Corporation (CDM Smith) in June 2014 calculated individual constituent and cumulative reasonable maximum exposure (RME) and central tendency exposure (CTE) cancer and non-cancer risks for current and reasonably anticipated future exposure scenarios and receptors, including adolescent and adult trespassers. The individual constituent RME cancer risks were less than United States Environmental Protection Agency (USEPA) target values for the receptors evaluated. The individual constituent RME non-cancer risks were greater than the USEPA target value (hazard index, HI) of 1 for adolescent and adult trespassers only. The non-cancer health hazard drivers are primarily PCBs for these receptors.

The results of the Baseline Ecological Risk Assessment (BERA) prepared by Integral Consulting, Inc. dated December 30, 2016, indicated that exposures to constituents of potential ecological concern in the environmental media at the Site do not pose an ecological concern for most of the evaluated receptors and that there is a low potential risk for short-tailed shrews and American robin. Therefore, this evaluation has focused on human health risk to assess the area where remediation is appropriate on the landfill portion of the Site.

Anticipated Future Use

The operations currently ongoing within the landfill (the Hunt Club, two landscape areas, and the storage of roll offs) will not continue beyond the completion of the soil remedial action at the Site. No future residential, industrial, commercial, recreational, or other use of the Site is anticipated.

DEVELOPMENT OF SITE-SPECIFIC RISK BASED CONCENTRATIONS

The adolescent trespasser receptor was used as the basis to develop the RBCs because it was the receptor with the highest potential health hazard. RBCs were calculated for PCBs, specifically non-dioxin-like PCBs, PCB toxic equivalent (TEQ), and dioxin TEQ; total xylenes; and antimony which account for approximately 90% of the cumulative health hazard.



Receptor: Adolescent Trespasser (Landfill)

	Exposure Point Concentration (mg/kg)	Soil Combined Dermal, Ingestion, and Inhalation HI	Percent of Cumulative HI	Site-Specific RBC (mg/kg)
Non-dioxin-like PCBs	57.42	3.3	57%	10
PCB TEQ	0.00042	0.7	12%	0.00007
Dioxin TEQ	0.00049	0.62	11%	0.00008
Total Xylenes	7,288	0.3	5%	1,300
Antimony	119	0.24	4%	21
	Cumulative HI:	5.8		

Notes:

1. Site-specific RBCs were calculated according to the following equation:

$$RBC = EPC \times \frac{\text{Target HI (1)}}{\text{Cumulative HI (5.8)}}$$

2. mg/kg = milligrams per kilogram

3. Site-specific RBCs are rounded to two significant figures or one significant figure if the value is 10 or less.

EXPOSURE POINT CONCENTRATIONS

To support the evaluation of the area to be remediated, an iterative approach was used to compare exposure point concentrations (EPCs) of PCBs, the primary risk driver and only constituent with an individual constituent hazard index above one (1), outside the proposed area to be remediated to the RBCs. Because the data sets for the PCB congener (non-dioxin-like PCBs and PCB TEQ) and dioxin (dioxin TEQ) were too small to support the evaluation of the impact of a cap on the EPC in readily accessible soil, Aroclor data (for which there was greater data density) was used. Total PCBs as the sum of Aroclors was the only constituent evaluated to support the determination of the area where remediation was appropriate.

This approach included removing the highest PCB concentrations from the data set in a step-wise manner, calculating the 95 percent upper confidence limit of the mean (95UCL) as the EPC for PCBs, comparing the EPC to the RBC, and continuing the process until the EPC was below the RBC.

USEPA's ProUCL version 5.1.002 was used to calculate the 95UCL as the EPC for PCBs (as the sum of Aroclors) for the shallow soils (0-2-foot depth interval) outside the area to be remediated as this would be accessible to an adolescent trespasser. The EPC for PCBs was calculated as 3.6 mg/kg. The ProUCL output is included as Attachment B-1 to this appendix.

CONCLUSIONS

Based on the iterative approach described previously, the area selected for remediation (Selected Area) is approximately 25 acres and is shown on Figure B-1. The EPC for PCBs in shallow soil outside the Selected Area is 3.6 mg/kg, and is below the RBC of 10 mg/kg at the Rolling Knolls Site; therefore, remediating the Selected Area will be protective of human health of potential receptors in the landfill portion of the Site.

REFERENCES

CDM Federal Programs Corporation, 2014. *Baseline Human Health Risk Assessment, Rolling Knolls Landfill Superfund Site, Chatham, New Jersey*. June.

Integral Consulting. 2016. *Baseline Ecological Risk Assessment, Rolling Knolls Landfill Superfund Site*. September.

TRC, 2017. *Reuse Assessment Report, Rolling Knolls Landfill Superfund Site, Chatham, New Jersey*. February.

USEPA, 2016. *ProUCL version 5.5.002*. June 20. <https://www.epa.gov/land-research/proucl-software>.

ProUCL Statistics for Data Sets with Non-Detects

User Selected Options

Date/Time of Computation ProUCL 5.111/9/2017 5:42:18 PM

From File aroclors - outside cap.xls

Full Precision OFF

Confidence Coefficient 95%

Number of Bootstrap Operations 2000

Aroclors

General Statistics

Total Number of Observations	110	Number of Distinct Observations	77
Number of Detects	81	Number of Non-Detects	29
Number of Distinct Detects	73	Number of Distinct Non-Detects	7
Minimum Detect	0.01	Minimum Non-Detect	0.03
Maximum Detect	29	Maximum Non-Detect	0.89
Variance Detects	22.68	Percent Non-Detects	26.36%
Mean Detects	3.67	SD Detects	4.762
Median Detects	2.29	CV Detects	1.297
Skewness Detects	3.024	Kurtosis Detects	12.2
Mean of Logged Detects	0.291	SD of Logged Detects	1.841

Normal GOF Test on Detects Only

Shapiro Wilk Test Statistic	0.706
5% Shapiro Wilk P Value	0
Lilliefors Test Statistic	0.221
5% Lilliefors Critical Value	0.0985

Normal GOF Test on Detected Observations Only

Detected Data Not Normal at 5% Significance Level

Lilliefors GOF Test

Detected Data Not Normal at 5% Significance Level

Detected Data Not Normal at 5% Significance Level

Kaplan-Meier (KM) Statistics using Normal Critical Values and other Nonparametric UCLs

KM Mean	2.708	KM Standard Error of Mean	0.419
KM SD	4.368	95% KM (BCA) UCL	3.439
95% KM (t) UCL	3.403	95% KM (Percentile Bootstrap) UCL	3.422
95% KM (z) UCL	3.397	95% KM Bootstrap t UCL	3.585
90% KM Chebyshev UCL	3.965	95% KM Chebyshev UCL	4.535
97.5% KM Chebyshev UCL	5.325	99% KM Chebyshev UCL	6.878

Gamma GOF Tests on Detected Observations Only

A-D Test Statistic	0.626
5% A-D Critical Value	0.807
K-S Test Statistic	0.0763
5% K-S Critical Value	0.104

Anderson-Darling GOF Test

Detected data appear Gamma Distributed at 5% Significance Level

Kolmogorov-Smirnov GOF

Detected data appear Gamma Distributed at 5% Significance Level

Detected data appear Gamma Distributed at 5% Significance Level

Gamma Statistics on Detected Data Only

k hat (MLE)	0.611	k star (bias corrected MLE)	0.596
Theta hat (MLE)	6.009	Theta star (bias corrected MLE)	6.154
nu hat (MLE)	98.94	nu star (bias corrected)	96.61
Mean (detects)	3.67		

Gamma ROS Statistics using Imputed Non-Detects

GROS may not be used when data set has > 50% NDs with many tied observations at multiple DLs

GROS may not be used when kstar of detects is small such as <1.0, especially when the sample size is small (e.g., <15-20)

For such situations, GROS method may yield incorrect values of UCLs and BTVs

This is especially true when the sample size is small.

For gamma distributed detected data, BTVs and UCLs may be computed using gamma distribution on KM estimates

Minimum	0.01	Mean	2.705
Maximum	29	Median	0.82
SD	4.389	CV	1.623
k hat (MLE)	0.339	k star (bias corrected MLE)	0.336
Theta hat (MLE)	7.983	Theta star (bias corrected MLE)	8.059
nu hat (MLE)	74.55	nu star (bias corrected)	73.85
Adjusted Level of Significance (β)	0.0478		
Approximate Chi Square Value (73.85, α)	55.06	Adjusted Chi Square Value (73.85, β)	54.85
95% Gamma Approximate UCL (use when $n \geq 50$)	3.628	95% Gamma Adjusted UCL (use when $n < 50$)	3.643

Estimates of Gamma Parameters using KM Estimates

Mean (KM)	2.708	SD (KM)	4.368
Variance (KM)	19.08	SE of Mean (KM)	0.419
k hat (KM)	0.384	k star (KM)	0.38
nu hat (KM)	84.58	nu star (KM)	83.61
theta hat (KM)	7.044	theta star (KM)	7.126
80% gamma percentile (KM)	4.343	90% gamma percentile (KM)	7.717
95% gamma percentile (KM)	11.45	99% gamma percentile (KM)	20.9

Gamma Kaplan-Meier (KM) Statistics

Approximate Chi Square Value (83.61, α)	63.53	Adjusted Chi Square Value (83.61, β)	63.3
95% Gamma Approximate KM-UCL (use when $n \geq 50$)	3.564	95% Gamma Adjusted KM-UCL (use when $n < 50$)	3.577

Lognormal GOF Test on Detected Observations Only

Shapiro Wilk Approximate Test Statistic 0.903

5% Shapiro Wilk P Value 1.0262E-6

Lilliefors Test Statistic 0.16

5% Lilliefors Critical Value 0.0985

Detected Data Not Lognormal at 5% Significance Level

Shapiro Wilk GOF Test

Detected Data Not Lognormal at 5% Significance Level

Lilliefors GOF Test

Detected Data Not Lognormal at 5% Significance Level

Lognormal ROS Statistics Using Imputed Non-Detects

Mean in Original Scale	2.716	Mean in Log Scale	-0.648
SD in Original Scale	4.383	SD in Log Scale	2.276

95% t UCL (assumes normality of ROS data)	3.409	95% Percentile Bootstrap UCL	3.426
95% BCA Bootstrap UCL	3.559	95% Bootstrap t UCL	3.639
95% H-UCL (Log ROS)	15.46		

Statistics using KM estimates on Logged Data and Assuming Lognormal Distribution

KM Mean (logged)	-0.877	KM Geo Mean	0.416
KM SD (logged)	2.519	95% Critical H Value (KM-Log)	3.972
KM Standard Error of Mean (logged)	0.25	95% H-UCL (KM -Log)	25.94
KM SD (logged)	2.519	95% Critical H Value (KM-Log)	3.972
KM Standard Error of Mean (logged)	0.25		

DL/2 Statistics

DL/2 Normal

Mean in Original Scale	2.712
SD in Original Scale	4.386
95% t UCL (Assumes normality)	3.405

DL/2 Log-Transformed

Mean in Log Scale	-0.8
SD in Log Scale	2.44
95% H-Stat UCL	21.8

DL/2 is not a recommended method, provided for comparisons and historical reasons

Nonparametric Distribution Free UCL Statistics

Detected Data appear Gamma Distributed at 5% Significance Level

Suggested UCL to Use

95% KM Approximate Gamma UCL	3.564
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Note: Suggestions regarding the selection of a 95% UCL are provided to help the user to select the most appropriate 95% UCL.

Recommendations are based upon data size, data distribution, and skewness.

These recommendations are based upon the results of the simulation studies summarized in Singh, Maichle, and Lee (2006).

However, simulations results will not cover all Real World data sets; for additional insight the user may want to consult a statistician.